

1 **BEHAVIOURAL STRATEGIES TOWARDS HUMAN DISTURBANCES EXPLAIN**
2 **INDIVIDUAL PERFORMANCE IN WOODLAND CARIBOU**

3

4 Authors ¹: Martin Leclerc ^a, Christian Dussault ^b and Martin-Hugues St-Laurent ^{a,*}

5

6 ^a Département de Biologie, Chimie et Géographie, Centre for Northern Studies & Centre for
7 Forest Research, Université du Québec à Rimouski, 300 Allée des Ursulines, Rimouski, Québec,
8 Canada, G5L 3A1.

9

10 ^b Ministère des Forêts, de la Faune et des Parcs du Québec, Direction de la faune terrestre et de
11 l'avifaune, 880 chemin Sainte-Foy, Québec, Québec, Canada, G1S 4X4.

12

13 * Corresponding Author: Martin-Hugues St-Laurent, Phone: (418) 723-1986 ext. 1538, Fax:
14 (418) 724-1849, E-mail: martin-hugues_st-laurent@uqar.ca

¹ Author Contributions: CD and MHS conceived and designed the project. ML, CD and MHS developed the methodology. ML analyzed the data. ML, CD and MHS wrote the manuscript.

15 **ABSTRACT** – Behavioural strategies may have important fitness, ecological and evolutionary
16 consequences. In woodland caribou, human disturbances are associated with higher predation
17 risk. Between 2004 and 2011, we investigated if habitat selection strategies of female caribou
18 towards disturbances influenced their calf’s survival in managed boreal forest with varying
19 intensities of human disturbances. Calf survival was 53% and 43% after 30 and 90 days
20 following birth, respectively and 52% of calves that died were killed by black bear. The
21 probability that a female lose its calf to predation was not influenced by habitat composition of
22 her annual home range, but decreased with an increase in proportion of open lichen woodland
23 within its calving home range. At the local scale, females that did not lose their calf displayed
24 stronger avoidance of high road density areas than females that lost their calf to predation.
25 Further, females that lost their calf to predation and that had a low proportion of ≤ 5 -year-old
26 cutovers within their calving home range were mostly observed in areas where these young
27 cutovers were locally absent. Also, females that lost their calf to predation and that had a high
28 proportion of ≤ 5 -year-old cutovers within their calving home range were mostly observed in
29 areas with a high local density of ≤ 5 -year-old cutovers. Our study demonstrates that we have to
30 account for human-induced disturbances at both local and regional scales in order to further
31 enhance effective caribou management plans. We demonstrate that disturbances not only impact
32 spatial distribution of individuals, but also their reproductive success.

33

34 **Key words:** Anthropogenic disturbances, calf survival, functional response, habitat selection,
35 reproductive success.

36 INTRODUCTION

37 Anthropogenic disturbances are widely spread across all ecosystems (Sanderson et al.
38 2002). Some species benefit from anthropogenic activities, however others, like those associated
39 with undisturbed habitats, are often negatively affected (Fisher and Wilkinson 2005). As outlined
40 by Johnson and St-Laurent (2011), anthropogenic infrastructure or disturbances may affect
41 animal physiology (Wikelski and Cooke 2006), behaviour (Blumstein et al. 2005), energetic
42 balance (Williams et al. 2006), survival (Phillips and Alldredge 2000), as well as populations and
43 communities (Addessi 1994).

44 The boreal forest, a biome representing 33% of the Earth's forest cover, is increasingly
45 impacted by forestry activities (Burton et al. 2003). The threatened woodland caribou *Rangifer*
46 *tarandus caribou* is recognized as being negatively impacted by both human-induced and natural
47 disturbances (Vors et al. 2007; Festa-Bianchet et al. 2011), and increased predation risk in highly
48 disturbed environments is considered the most important proximate limiting factor explaining the
49 widespread population declines (Wittmer et al. 2007; Festa-Bianchet et al. 2011). Forestry
50 activities benefit wolf *Canis lupus* (Seip 1991), the main predator of adult caribou, and black bear
51 *Ursus americanus* which is another potential predator of caribou calves (Gustine et al. 2006;
52 Pinard et al. 2012). Although adult female survival is the most important determinant of
53 population dynamics, calf survival is highly variable among years and populations (Seip and
54 Cichowski 1996; Gustine et al. 2006; Pinard et al. 2012) and it also has a great influence on
55 population dynamics (Gaillard et al. 2000; Raithel et al. 2007).

56 During the calving period, female caribou behavioural adjustments are aimed at reducing
57 predation risk for their calf (Gustine et al. 2006; Pinard et al. 2012), and their habitat selection is
58 oriented toward old-growth coniferous forests (Lantin et al. 2003; Mahoney and Virgl 2003),
59 open lichen woodlands and peatlands (McLoughlin et al. 2005; Hins et al. 2009). Females also

60 select areas located at high elevations or in rugged terrain, while avoiding cutovers, regenerating
61 areas and roads (Leclerc et al. 2012; Leblond et al. 2013). Caribou attempt to isolate themselves
62 spatially from predators, and their habitat selection pattern appears to be effective in avoiding
63 predation by wolves (James et al. 2004). However, calves also suffer from black bear predation in
64 landscapes where the human footprint is extensive (Mahoney and Virgl 2003; Pinard et al. 2012).
65 Some authors have suggested that wolf-avoidance strategies displayed by caribou could result in
66 increased bear predation (Faille et al. 2010; St-Laurent and Dussault 2012), a potentially
67 maladaptive behaviour due to recent increases in bear densities across the caribou range.

68 During the last decade, an increasing number of studies have investigated differences in
69 behaviour (Réale et al. 2010) which can occur between and within populations of the same
70 species (Wilson 1998). For example, individuals may express different habitat selection (Mabille
71 et al. 2012) or movement patterns while foraging (Austin et al. 2004) or different intensities of
72 boldness or shyness (Wilson et al. 1994; Réale and Festa-Bianchet 2003). Because different
73 behavioural strategies may lead to different survival probabilities or reproductive rates,
74 behavioural strategies may have fitness, ecological and evolutionary consequences (Sih et al.
75 2004; Smith and Blumstein 2008; Réale et al. 2010). Therefore it is important to consider the
76 range of behavioural strategies displayed by individuals when assessing the impacts of human-
77 induced disturbances on wildlife, particularly in caribou, as human disturbances are associated
78 with higher predation risk for that species (Courtois et al. 2008; Whittington et al. 2011).

79 The study of functional responses in habitat selection may help to highlight the different
80 behavioural strategies within a population. A functional response in habitat selection is defined as
81 a change in the selection of a habitat attribute as a function of its availability or the availability of
82 other habitat attributes (Mysterud and Ims 1998; Hebblewhite and Merrill 2008). Functional
83 responses in habitat selection have been reported to occur in large ungulates such as moose *Alces*

84 *alces* (Mabille et al. 2012; Beyer et al. 2013), red deer *Cervus elaphus* (Godvik et al. 2009) and
85 caribou (Hansen et al. 2009; Moreau et al. 2012). The study of functional responses towards
86 human disturbances is important for managers as the impacts of disturbances can vary according
87 to their local and regional density. Furthermore, functional responses in habitat selection that are
88 triggered by human disturbances might have important fitness implications and ultimately alter
89 animal's adaptive value. Despite such important potential consequences on population dynamics,
90 few studies have tried to link functional responses in behaviour to fitness (except Dussault et al.
91 2012). In this study, we used GPS collars to track female caribou in managed landscapes with
92 varying intensities of human disturbances in the boreal forest of Québec, Canada. We first
93 evaluated calf survival and then investigated if habitat selection strategies of females towards
94 major human disturbances influenced their reproductive success. Further, we examined if
95 functional responses in habitat selection of adult female caribou towards human-induced
96 disturbances could explain the fate of their calf.

97

98 **MATERIALS AND METHODS**

99 **Study area**

100 The study area (Supplemental Figure S1) included two caribou ranges, i.e., Charlevoix
101 (5086 km²; 47°40' N, 71°15' W) and Saguenay–Lac-St-Jean (26 686 km²; 48°28'–50°59' N,
102 69°59'–72°15' W), both typical of the boreal forest but contrasted in terms of human footprint. In
103 both regions, the landscape was dominated by coniferous stands composed of black spruce *Picea*
104 *mariana* and balsam fir *Abies balsamea* with few mixed or deciduous stands composed of white
105 birch *Betula papyrifera*, trembling aspen *Populus tremuloides*, and maples *Acer spp.* Topography
106 was characterized by low rolling relief ranging between 250 and 900 m in Saguenay–Lac-St-
107 Jean, and between 500 and 1000 m in Charlevoix (Robitaille and Saucier 1998). Mean

108 precipitation was 1500 mm y⁻¹ in Charlevoix and 1200 mm y⁻¹ in Saguenay–Lac-St-Jean, and
109 snow accumulations could reach >3 m (Robitaille and Saucier 1998). The Charlevoix study
110 region encompassed 3 National Parks (Grands-Jardins, Jacques-Cartier, and Hautes-Gorges-de-
111 la-Rivière-Malbaie) where forest harvesting was prohibited; the study region also includes a large
112 part of the Laurentides Wildlife Reserve where forestry activities were allowed, such as in
113 Saguenay–Lac-St-Jean. Human-disturbed stands (cutovers of varying ages only) occupied ca.
114 44% of the caribou range in Charlevoix and 32% in Saguenay–Lac-St-Jean. When applying a
115 500-m buffer to anthropogenic disturbances (cutovers < 50 year-old and roads) according to the
116 Environment Canada (2011) model, the proportion of each caribou range under the influence of
117 human-induced disturbances reached ca. 99% in Charlevoix and 77% in Saguenay–Lac-St-Jean.
118 Other large mammals inhabiting the study area were moose, black bear, and grey wolf. White-
119 tailed deer *Odocoileus virginianus* were also occasionally seen in Charlevoix.

120

121 **Capture and monitoring**

122 Between 2004 and 2011, we captured female caribou using a net-gun fired from a
123 helicopter (Potvin and Breton 1988), and equipped them with GPS collars (model 2200L or
124 3300L from Lotek, Newmarket, Ontario, Canada or model TGW4600 from Telonics, Mesa, AZ,
125 USA). We programmed GPS collars to attempt location fixes every 1, 2, 3 or 4 hours depending
126 upon the collar model and study site.

127 From 2004 to 2007 (Charlevoix) and 2009 to 2011 (Saguenay–Lac-St-Jean), we captured as
128 many calves as possible from females equipped with GPS collars. To do so, we conducted
129 telemetry flights spaced < 3 days apart during the calving period (21-May to 20-June) to locate
130 collared females and visually confirm if a newborn calf was absent or present. When a calf was
131 first detected, we evaluated if we could capture it immediately (calf standing firmly or moving

132 with its mother) or delayed capture to the next day (wet fur, lying down, low mobility). We
133 captured each calf by hand wearing latex gloves, determined its sex, and fitted it with a VHF
134 expandable collar equipped with a mortality sensor (model LMRT-3 from Lotek, Newmarket,
135 Ontario, Canada or model M2510B from Advanced Telemetry Systems, Isanti, MN, USA). We
136 have no evidence that the capture of a calf affected its survival probability as our survival
137 estimates were consistent with the cow : calf ratio observed in the population during aerial
138 surveys (Courtois et al. 2007; Pinard et al. 2012). All capture and handling procedures were
139 approved by the Animal Welfare Committee of the Université du Québec à Rimouski
140 (certificates #36-08-67 and #27-07-53) and of the Ministère du Développement durable, de
141 l'Environnement, de la Faune et des Parcs du Québec (certificate #07-00-02).

142 We monitored calf survival by flying over the study area to detect any VHF mortality
143 signals at least twice a week from birthing to the end of June (flights being on average 1.8 ± 1.5
144 [SD] days apart), and every 2 weeks thereafter until late August (flights being on average $10.8 \pm$
145 6.7 [SD] days apart), and finally once per month in September and October. When a mortality
146 signal was detected, we located the collar within 72 hours of detection and investigated the
147 carcass and its surroundings to determine cause of mortality. We considered the presence of tooth
148 marks, blood or disarticulated, dispersed or crushed bones as evidence of predation, and we
149 determined the predator species with evidence such as faeces, tracks, hairs, and scratching signs
150 (Pinard et al. 2012). We used the date of telemetry flights, evidence left at the mortality site, and
151 unusually long movements of a calf's mother, based on their GPS locations, to estimate the
152 mortality date more precisely.

153

154 **Spatial analyses**

155 We used 1 : 20 000 digital ecoforest maps, published by the Quebec government, to
156 describe caribou habitat. We updated these maps annually to include new habitat modifications
157 resulting from forestry practices and natural disturbances. Minimum mapping unit size was 4 ha
158 for forested polygons and 2 ha for non-forested areas (e.g., water bodies). We combined polygons
159 available on ecoforest maps into 10 habitat types based on caribou ecology: (1) ‘Coniferous’ and
160 (2) ‘Mixed and deciduous’ stands included stands with dominant coniferous or mixed and
161 deciduous tree strata ≥ 50 -year-old, respectively; (3) ‘Open lichen woodlands’ referred to
162 coniferous forest with terrestrial lichens and low tree density; (4) ‘Peatlands’ were poorly drained
163 open areas (mainly bogs and fens); cutovers were divided using time since disturbance, resulting
164 in (5) ‘ ≤ 5 -year-old cutovers’, (6) ‘6-20-year-old cutovers’, (7) ‘20-40-year-old cutovers’ and (8)
165 ‘Open no regeneration’ which referred to areas originating from a natural disturbance but with no
166 established tree and shrub strata; (9) ‘Water bodies’ and (10) ‘Others’, which primarily included
167 non-forested areas. We also created a digital elevation model (cell size: 30×30 m) using 1 : 20
168 000 topographic maps.

169 We investigated the link between habitat selection of adult females and the fate of their calf
170 at three spatial scales: 1) the annual home range scale, 2) the calving home range scale, and 3) the
171 local scale. To do so, we subsampled the GPS telemetry database to get location intervals of 3 or
172 4 h. At both the annual and calving home range scales, we compared the home range composition
173 between females that lost and those that did not lose their calf. We defined home ranges using
174 100% minimum convex polygon (MCP).

175 At the local scale, we contrasted habitat use and availability by comparing locations of each
176 female caribou to an equal number of locations randomly distributed within their annual home
177 range (3rd order of selection, *sensu* Johnson 1980). We defined home ranges using 100% MCP.
178 We determined elevation, and calculated the proportion of each undisturbed and disturbed habitat

179 type, as well as road density within an 816-m radius circular buffer centred on each GPS and
180 random location. This allowed us to consider the influence of the surrounding environment on
181 habitat selection at the local scale (Leblond et al. 2011). We used an 816-m buffer size because it
182 represented the median daily distance travelled by females during the calving period. We
183 conducted all spatial analyses using ArcGIS 10.0 (ESRI Inc., Redlands, California, USA).

184

185 **Statistical analyses**

186 We assessed calf survival rate using a Cox Proportional Hazards regression model (Cox
187 1972; McLoughlin et al. 2005) and tested the effect of year, sex, and date of birth on calf survival
188 using the “Survival” library in R 2.15.1 (R Development Core Team 2012). For this survival
189 analysis, we only used calves from the Saguenay–Lac-St-Jean region as calf survival curves for
190 the Charlevoix region were published by Pinard et al. (2012). We nevertheless graphed the
191 survival functions of both regions.

192 For the habitat selection analyses, we used GPS locations of females from Saguenay–Lac-
193 St-Jean and Charlevoix for which we also had calf survival data. Females that lost their calf from
194 a cause other than predation were removed from our analysis, as our focus was on mortality by
195 predation and not total mortality (McLoughlin et al. 2005; Dussault et al. 2012). We further
196 removed 3 other females from our habitat selection analyses because we did not retrieve their
197 GPS collar ($n = 2$) or we did not have ecoforest maps for their location ($n = 1$).

198 At the annual and calving home range scales, we used mixed effects logistic regressions to
199 determine the influence of the annual and calving home range composition on female
200 reproductive success. We used calf fate (0 = alive, 1 = dead from predation) as the dependent
201 variable and mean elevation, proportion of each undisturbed and disturbed habitat type, and road
202 density within the female annual and calving home ranges as independent variables. We also

203 considered the region (i.e., Saguenay–Lac-St-Jean or Charlevoix) as a random effect. We
204 performed model selection and evaluated different alternative hypotheses (3 candidate models;
205 see Table 1) using Akaike’s Information Criterion (AIC). Further, we tested if the best-supported
206 model was statistically different from the null model using the likelihood ratio test.

207 At the local scale, we aimed to highlight the habitat selection strategies of females, not the
208 differences between the mortality site and previous locations (which would have required using
209 Cox models; Cox 1972). Therefore, in order to compare habitat selection between females that
210 lost their calf to predation and females for which the calf survived throughout the study period,
211 we used Resource Selection Functions (RSFs; Manly et al. 2002). Following Dussault et al.
212 (2012), we randomly matched a female for which the calf was killed by a predator to a female
213 whose calf survived. This approach did not yield a true pairing in a statistical sense since there
214 was no link between females within a pair, but it allowed us to perform balanced regression
215 models, i.e. with an equal number of females that lost and that did not lose their calf. For each
216 pairing, we restricted the dataset of each female with a calf that survived to the same number of
217 days as the one that died, the counting starting at calving. Doing so allowed us to compare habitat
218 selection of females while controlling for calf age, and prevented us from detecting differences
219 due only to increased calf mobility or environmental changes which both changed along with calf
220 age (Dussault et al. 2012). We pooled the resulting datasets and used this new file to conduct
221 RSFs. We used mixed logistic regressions to compare habitat characteristics at recorded female
222 locations (use) to those at random locations (available) and considered the calf nested in female,
223 nested in region, as a random effect. We repeated this process 999 times, each time with a
224 different pairing of calves that died and that survived. In each iteration, we used the new database
225 to calculate RSFs and evaluate different candidate models (Table 1) using AIC. In candidate
226 models 4 to 12, we added the interaction between calf fate and habitat type covariates in order to

227 evaluate if habitat selection strategies of females could be linked to the fate of their calf. Further,
228 in models 7 to 12, we added triple interactions to explore if functional responses in habitat
229 selection could explain calf fate. Thus, for each candidate model, we obtained 1000 coefficient
230 estimates that we used to calculate the mean coefficient and associated 95% CIs (determined as
231 the 2.5 and 97.5 percentiles). Prior to all habitat selection analyses, we assessed multicollinearity
232 between independent variables using the variance inflation factor (VIF; threshold of 5; Graham
233 2003) and consequently removed the ‘% of coniferous stands’ variable from the candidate
234 models. We conducted all statistical analyses using R 2.15.1.

235

236 **RESULTS**

237 Between 2009 and 2011, we captured 30 calves in Saguenay–Lac-St-Jean. In this region,
238 mean calf survival rate was 53% and 43% after 30 and 90 days following birth, respectively
239 (Table 2, Fig. 1). Nine of the 17 (53%) calves that died were killed by black bear, the most
240 important mortality agent, and no calf died of wolf predation. Four calves (13%) died from an
241 unknown natural cause (Table 2). Most calf mortalities occurred during the first month of life and
242 survival rate stabilized after 90 days (Fig. 1). Survival rate did not vary with calf sex ($\chi^2 = 0.3$; df
243 $= 1$; $P = 0.559$), year ($\chi^2 = 1.5$; $df = 2$; $P = 0.471$), and birth date ($\chi^2 = 22.2$; $df = 16$; $P = 0.136$).
244 In Charlevoix, mean calf survival rate was 47% after 90 days and black bear was also the primary
245 mortality agent (see Pinard et al. 2012 for more information).

246 We conducted the habitat selection analyses using GPS locations of 22 mothers from the
247 Saguenay–Lac-St-Jean region, from which 11 calves survived and 11 died from predation, and 35
248 mothers from Charlevoix, from which 16 calves survived and 19 died from predation. At the two
249 largest spatial scales, calf fate did not depend on the reaction of their mother toward human
250 disturbances. At the annual home range scale, the best-supported model describing calf fate only

251 included undisturbed habitat types (Table 3). However, this model did not differ from the null
252 model ($\chi^2 = 2.09$; $df = 4$; $P = 0.718$). At the calving home range scale, the best-supported model
253 also included undisturbed habitat types only (Table 3) but in this case, it was statistically better
254 than the null model ($\chi^2 = 10.36$; $df = 4$; $P = 0.035$). The probability that a female lose its calf to
255 predation decreased as the proportion of open lichen woodlands in its calving home range
256 increased (Table 4).

257 At the local scale, the best-supported model revealed that calf fate was linked to habitat
258 selection strategies of adult females toward human disturbances (Table 3). Three major results
259 emerged from this analysis. First, all females avoided areas of high road density, but females that
260 did not lose their calf displayed stronger avoidance of high road density areas than females that
261 lost their calf to predation (Table 4). Second, females that did not lose their calf to predation
262 displayed stronger selection of mixed and deciduous stands (Table 4). Finally, calf fate depended
263 on the combined local (within the 816-m buffer) density of ≤ 5 -year-old cutovers and on the
264 proportion of ≤ 5 -year-old cutovers within the calving home range (Table 4, Fig. 2). This
265 functional response towards ≤ 5 -year-old cutovers highlights two different habitat selection
266 strategies that resulted in the same calf fate. Females that lost their calf to predation and that had
267 a low proportion of ≤ 5 -year-old cutovers within their calving home range were mostly observed
268 in areas where ≤ 5 -year-old cutovers were locally absent (Fig. 2). Also, females that lost their calf
269 to predation and that had a high proportion of ≤ 5 -year-old cutovers within their calving home
270 range were mostly observed in areas with a high local density of ≤ 5 -year-old cutovers (Fig. 2).

271

272 **DISCUSSION**

273 Our study demonstrates that differences in behavioural strategy towards human-induced
274 disturbances led to different calf fate outcomes in a large ungulate, the threatened woodland

275 caribou. Different habitat selection strategies displayed by females toward ≤ 5 -year-old cutovers
276 and roads resulted in different outcomes in reproductive success. At the local scale, all human
277 disturbances induced female caribou behavioural changes, but not all had consequences on the
278 fate of their calf. Female's habitat selection patterns at the largest spatial scale (annual home
279 range) did not influence the probability that her calf died from predation, suggesting that females
280 avoided predation risk at smaller spatial and temporal scales. Nevertheless, including open lichen
281 woodlands, an undisturbed habitat type known to be important to caribou for spatially
282 segregating them from predators and alternative prey (Hins et al. 2009), in a calving home range
283 appears to increase a calf's survival probability.

284 We further demonstrated that the functional response in habitat selection by females
285 towards ≤ 5 -year-old cutovers explained some variation in calf fate. During the lactation period,
286 the energetic requirements of females double (Chan-McLoed et al. 1994) so that fine scale habitat
287 selection might be oriented towards habitat types with higher food availability (Lantin et al.
288 2003; Carr et al. 2007). The use of ≤ 5 -year-old cutovers, a highly productive habitat type, likely
289 resulted in more frequent foraging opportunities for female caribou (Bergerud 1972; Bock and
290 Van Rees 2002). We suggest two plausible hypotheses to explain the adaptive use of ≤ 5 -year-old
291 cutovers by female caribou after calving. Higher food availability may result in 1) higher milk
292 production and in 2) more time available for maternal care, both of which should promote calf
293 growth (White 1983; Rognmo et al. 1983), and shorten the period during which calves are highly
294 vulnerable to predators. However, when the proportion of ≤ 5 -year-old cutovers within the
295 landscape reaches higher levels, the positive effect of increased food availability is superseded by
296 the negative effect of increased predation risk. Disturbed habitat types, such as ≤ 5 -year-old
297 cutovers, are often associated with higher predation risk (Wittmer et al. 2007; Festa-Bianchet et
298 al. 2011) and the proportion of disturbed habitat types within caribou habitat has been negatively

299 correlated with calf recruitment (Environment Canada 2011). The increased proportion of
300 disturbed habitat types triggers numerical (Seip 1991) and functional (Houle et al. 2010)
301 responses of caribou predators. Indeed, wolves are known to increase their selection of recent
302 cutovers when these areas are more abundant in the landscape (Houle et al. 2010). A similar
303 functional response towards ≤ 5 -year-old cutovers could also exist for other predators of caribou
304 calves, such as black bear.

305 Calf survival was primarily influenced by black bear predation in Saguenay–Lac-St-Jean
306 (Pinard et al. 2012). Wolves killed 5% of calves in Charlevoix and none in Saguenay–Lac-St-
307 Jean, suggesting that the wolf-avoidance strategy used by female caribou during the calving
308 period is currently effective in these areas. This wolf-avoidance strategy mainly consists of
309 selecting higher elevations (Leclerc et al. 2012; Pinard et al. 2012) as wolves usually use lower
310 elevations to move through the landscape (Whittington et al. 2011; Lesmerises et al. 2012).
311 However, by avoiding wolves, the predator with which caribou co-evolved, caribou could be at
312 increased predation risk by black bear, as the latter also selected higher elevations during spring
313 (Mosnier et al. 2008). We also hypothesize that females might have difficulties in assessing calf
314 predation risk by bears because they rarely attack adult caribou (Ballard 1994). Although black
315 bears killed the most calves in our system, Bastille-Rousseau et al. (2011) suggested that bears
316 prey only opportunistically on caribou calves during their frequent movements between food-rich
317 habitat patches such as regenerating stands.

318 We also demonstrated that linear infrastructure negatively influenced caribou behaviour
319 and calf survival. All females avoided areas with high road density, but females that avoided
320 these linear features more strongly were less likely to lose their calf to predation. Caribou
321 avoidance of roads has been demonstrated to be a means of avoiding wolves (James and Stuart-
322 Smith 2000; Leclerc et al. 2012) because roads are associated with higher predation risk by wolf

323 (James and Stuart-Smith 2000; Whittington et al. 2011). Roads may also increase black bear
324 predation risk as roadsides are highly productive environments that are selected by bear during
325 spring (Bastille-Rousseau et al. 2011).

326 Females that lost and those that did not lose their calf to predation displayed different
327 habitat selection strategies towards young cutovers. Some females selected ≤ 5 -year-old cutovers
328 even if it resulted in higher predation risk for their calf. We suggest that such behavioural
329 response could result in an ecological trap when an individual continues to frequent a formerly
330 suitable habitat patch that was modified by human activities (Schlaepfer et al. 2002). This
331 maladaptive behaviour, although expressed only by some females, could jeopardize caribou
332 recruitment particularly in highly managed boreal forest regions where black bear densities are
333 high, and could at least partially explain the mechanism linking habitat alteration with the global
334 decline of woodland caribou.

335

336 **ACKNOWLEDGEMENTS**

337 We thank B. Baillargeon, C. Bourgeois, L. Breton, L. Coulombe, R. Courtois, Cl. Dussault,
338 J.-G. Frenette, S. Gravel, D. Grenier, R. Lavoie, D. Lacasse, M. Poulin and S. St-Onge for
339 caribou captures. We also thank J.-P. Ouellet for his scientific contribution, A. Caron and M.
340 Mazerolle for statistical advices, and J. Bêty, C. Johnson, and P. and M. Fast for useful comments
341 on earlier versions of the manuscript. This project was funded by the Fonds de recherche du
342 Québec – Nature et technologies, the Fonds de recherche forestière du Saguenay–Lac-St-Jean,
343 the Natural Sciences and Engineering Research Council of Canada (Discovery Grant to M.-H. St-
344 Laurent), the Ministère du Développement durable, de l'Environnement, de la Faune et des Parcs
345 du Québec, the Conseil de l'industrie forestière du Québec, the Fédération canadienne de la
346 faune, the Fondation de la faune du Québec, the World Wildlife Fund for Nature, Produits

347 forestier Résolu Inc., and the Université du Québec à Rimouski. We also thank the Essipit First
348 Nation for providing access to their caribou telemetry data, via the Aboriginal Funds for Species
349 at Risk (Environment Canada).

350 **REFERENCES**

- 351 Addressi L (1994) Human disturbance and long-term changes on a rocky intertidal community.
352 *Ecol Appl* 4:786-797
- 353 Austin D, Bowen WD, McMillan JI (2004) Intraspecific variation in movement patterns:
354 Modeling individual behaviour in a large marine predator. *Oikos* 105:15-30 doi:
355 10.1111/j.0030-1299.1999.12730.x
- 356 Ballard WB (1994) Effects of black bear predation on caribou - A review. *Alces* 30:25-35
- 357 Bastille-Rousseau G, Fortin D, Dussault C, Courtois R, Ouellet J-P (2011) Foraging strategies by
358 omnivores: are black bears actively searching for ungulate neonates or are they simply
359 opportunistic predators? *Ecography* 34:588-596 doi: 10.1111/j.1600-0587.2010.06517.x
- 360 Bergerud AT (1972) Food habits of Newfoundland caribou. *J Wildl Manage* 36:913-923
- 361 Beyer HL, Ung R, Murray DL, Fortin M-J (2013) Functional responses, seasonal variation and
362 thresholds in behavioural responses of moose to road density. *J Appl Ecol* 50:286-294 doi:
363 10.1111/1365-2664.12042
- 364 Blumstein DT, Fernandez-Juricic E, Zollner PS, Garity SC (2005) Inter-specific variation in
365 avian responses to human disturbance. *J Appl Ecol* 42:943-953 doi: 10.1111/j.1365-
366 2664.2005.01071.x
- 367 Bock MD, Van Rees KCJ (2002) Forest harvesting impacts on soil properties and vegetation
368 communities in the Northwest Territories. *Can J For Res* 32:713-724 doi: 10.1139/X02-014
- 369 Burton PJ, Messier C, Weetman GF, Prepas EE, Adamowicz WL, Tittler R (2003) The current
370 state of boreal forestry and the drive for change. In: Burton PJ, Messier, C, Smith DW,
371 Adamowicz WL (eds) *Towards sustainable management of the boreal forest*. NRC
372 Research Press, pp 1-40

373 Carr NL, Rodgers AR, Walshe S (2007) Caribou nursery site habitat characteristics in two
374 northern Ontario parks. *Rangifer* 17:167-179 doi: <http://dx.doi.org/10.7557/2.27.4.343>

375 Chan-McLoed ACA, White RG, Holleman DF (1994) Effects of protein and energy-intake, body
376 condition, and season on nutrient partitioning and milk production in caribou and reindeer.
377 *Can J Zool* 72:938-947

378 Courtois R, Ouellet J-P, Breton L, Gingras A, Dussault C (2007) Effects of forest disturbance on
379 density, space use, and mortality of woodland caribou. *Ecoscience* 14:491-498

380 Courtois R, Gingras A, Fortin D, Sebbane A, Rochette B, Breton L (2008) Demographic and
381 behavioural response of woodland caribou to forest harvesting. *Can J For Res* 38:2837-
382 2849 doi: 10.1139/X08-119

383 Cox DR (1972) Regression models and life-tables. *J Roy Stat Soc B* 34:187-220

384 Dussault C, Pinard V, Ouellet J-P, Courtois R., Fortin D (2012) Avoidance of roads and selection
385 for recent cutovers by threatened caribou: fitness-rewarding or maladaptive behaviour?
386 *Proc R Soc B* 279:4481-4488 doi:10.1098/rspb.2012.1700

387 Environment Canada (2011) Scientific assessment to inform the identification of critical habitat
388 for woodland caribou (*Rangifer tarandus caribou*), Boreal population, in Canada: 2011
389 update. Environment Canada, Ottawa

390 Faille G, Dussault C, Ouellet J-P, Fortin D, Courtois R, St-Laurent M-H, Dussault C (2010)
391 Range fidelity: The missing link between caribou decline and habitat alteration? *Biol Cons*
392 143:2840-2850 doi:10.1016/j.biocon.2010.08.001

393 Festa-Bianchet M, Ray JC, Côté SD, Gunn A (2011) Conservation of caribou (*Rangifer*
394 *tarandus*) in Canada: an uncertain future. *Can J Zool* 89:419-434 doi:10.1139/Z11-025

395 Fischer JT, Wilkinson L (2005) The response of mammals to forest fire and timber harvest in the
396 North American boreal forest. *Mammal Rev* 35:51-81 doi: 10.1111/j.1365-
397 2907.2005.00053.x

398 Gaillard J-M, Festa-Bianchet M, Yoccoz NG, Loison A, Toigo C (2000) Temporal variation in
399 fitness components and population dynamics of large herbivores. *Ann Rev Ecol Syst*
400 31:367-393

401 Godvik IMR, Loe LE, Vik JO, Veiberg V, Langvatn R, Mysterud A (2009) Temporal scales,
402 trade-offs, and functional responses in red deer habitat selection. *Ecology* 90:699-710 doi:
403 <http://dx.doi.org/10.1890/08-0576.1>

404 Graham MH (2003) Confronting multicollinearity in ecological multiple regression. *Ecology*
405 84:2809-2815 doi: <http://dx.doi.org/10.1890/02-3114>

406 Gustine DD, Parker KL, Lay RJ, Gillingham MP, Heard DC (2006) Calf survival of woodland
407 caribou in a multi-predator ecosystem. *Wildl Monogr* 165:1-32 doi:
408 [http://dx.doi.org/10.2193/0084-0173\(2006\)165\[1:CSOWCI\]2.0.CO;2](http://dx.doi.org/10.2193/0084-0173(2006)165[1:CSOWCI]2.0.CO;2)

409 Hansen BB, Herfindal I, Aanes R, Saether B-E, Henriksen S (2009) Functional response in
410 habitat selection and the tradeoffs between foraging niche components in a large herbivore.
411 *Oikos* 118:859-872 doi: 10.1111/j.1600-0706.2009.17098.x

412 Hebblewhite M, Merrill E (2008) Modelling wildlife–human relationships for social species with
413 mixed-effects resource selection models. *J Appl Ecol* 45:834-844 doi: 10.1111/j.1365-
414 2664.2008.01466.x

415 Hins C, Ouellet J-P, Dussault C, St-Laurent M-H (2009) Habitat selection by forest-dwelling
416 caribou in managed boreal forest of eastern Canada: Evidence of a landscape configuration
417 effect. *For Ecol Manage* 257:636-643 doi: <http://dx.doi.org/10.1016/j.foreco.2008.09.049>

418 Houle M, Fortin D, Dussault C, Courtois R, Ouellet J-P (2010) Cumulative effects of forestry on
419 habitat use by gray wolf (*Canis lupus*) in the boreal forest. *Landsc Ecol* 25:419-433 doi:
420 10.1007/s10980-009-9420-2

421 James ARC, Stuart-Smith AK (2000) Distribution of caribou and wolves in relation to linear
422 corridors. *J Wildl Manage* 64:154-159

423 James ARC, Boutin S, Hebert DM, Blair Rippin A (2004) Spatial separation of caribou from
424 moose and its relation to predation by wolves. *J Wildl Manage* 68:799-809 doi:
425 [http://dx.doi.org/10.2193/0022-541X\(2004\)068\[0799:SSOCFM\]2.0.CO;2](http://dx.doi.org/10.2193/0022-541X(2004)068[0799:SSOCFM]2.0.CO;2)

426 Johnson DH (1980) The comparison of usage and availability measurements for evaluating
427 resource preference. *Ecology* 61:65-71

428 Johnson CJ, St-Laurent M-H (2011) Unifying framework for understanding impacts of human
429 developments on wildlife. In: Naugle DE (ed) *Energy development and wildlife*
430 *conservation in western North America*. Island Press, Washington, pp 27-54

431 Lantin É, Drapeau P, Paré M, Bergeron Y (2003) Preliminary assessment of habitat
432 characteristics of woodland caribou calving areas in the Claybelt region of Québec and
433 Ontario, Canada. *Rangifer* 14:247-254 doi: <http://dx.doi.org/10.7557/2.23.5.1708>

434 Leblond M, Frair J, Fortin D, Dussault C, Ouellet J-P, Courtois R (2011) Assessing the influence
435 of resource covariates at multiple spatial scales: an application to forest-dwelling caribou
436 faced with intensive human activity. *Landsc Ecol* 26:1433-1446 doi: 10.1007/s10980-011-
437 9647-6

438 Leblond M, Dussault C, Ouellet J-P (2013) Avoidance of roads by large herbivores and its
439 relation to disturbance intensity. *J Zool* 289:32-40 doi: 10.1111/j.1469-7998.2012.00959.x

440 Leclerc M, Dussault C, St-Laurent M-H (2012) Multiscale assessment of the impacts of roads
441 and cutovers on calving site selection in woodland caribou. For *Ecol Manage* 286:59-65
442 doi: <http://dx.doi.org/10.1016/j.foreco.2012.09.010>

443 Lesmerises F, Dussault C, St-Laurent M-H (2012) Wolf habitat selection is shape by human
444 activities in a highly managed boreal forest. For *Ecol Manage* 276:125-131 doi:
445 <http://dx.doi.org/10.1016/j.foreco.2012.03.025>

446 Mabile G, Dussault C, Ouellet J-P, Laurian C (2012) Linking trade-offs in habitat selection with
447 the occurrence of functional responses for moose living in two nearby study areas.
448 *Oecologia* 170:965-977 doi: 10.1007/s00442-012-2382-0

449 Mahoney SP, Virgl JA (2003) Habitat selection and demography of a nonmigratory woodland
450 caribou population in Newfoundland. *Can J Zool* 81:321-334 doi: 10.1139/z02-239

451 Manly BFJ, McDonald LL, Thomas DL, McDonald TL, Erickson WP (2002) Resource selection
452 by animals: Statistical design and analysis for field studies, 2nd edition. Kluwer Academic
453 Publishers, Dordrecht

454 McLoughlin PD, Dunford JS, Boutin S (2005) Relating predation mortality to broad-scale habitat
455 selection. *J Anim Ecol* 74:701-707 doi: 10.1111/j.1365-2656.2005.00967.x

456 Moreau G, Fortin D, Couturier S, Duchesne T (2012) Multi-level functional responses for
457 wildlife conservation: the case of threatened caribou in managed boreal forest. *J Appl Ecol*
458 49:611-620 doi: 10.1111/j.1365-2664.2012.02134.x

459 Mosnier A, Ouellet J-P, Courtois R (2008) Black bear adaptation to low productivity in the
460 boreal forest. *Ecoscience* 15:485-497 doi: 10.2980/15-4-3100

461 Mysterud A, Ims RA (1998) Functional responses in habitat use: availability influences relative
462 use in trade-off situations. *Ecology* 79:1435-1441 doi: [http://dx.doi.org/10.1890/0012-](http://dx.doi.org/10.1890/0012-9658(1998)079[1435:FRIHUA]2.0.CO;2)
463 [9658\(1998\)079\[1435:FRIHUA\]2.0.CO;2](http://dx.doi.org/10.1890/0012-9658(1998)079[1435:FRIHUA]2.0.CO;2)

464 Phillips GE, Alldredge AW (2000) Reproductive success of elk following disturbance by humans
465 during calving season. *J Wildl Manage* 64:521-530

466 Pinard V, Dussault C, Ouellet J-P, Fortin D, Courtois R (2012) Calving rate, calf survival rate,
467 and habitat selection of forest-dwelling caribou in a highly managed landscape. *J Wildl*
468 *Manage* 76:189-199 doi: 10.1002/jwmg.217

469 Potvin F, Breton L (1988) Use of a net gun for capturing white-tailed deer, *Odocoileus*
470 *virginianus*, on Anticosti Island, Québec. *Can Field Nat* 102:697-700

471 R Development Core Team (2012) R: A language and environment for statistical computing. R
472 foundation for statistical computing, Austria

473 Raitheil JD, Kauffman MJ, Pletscher DH (2007) Impact of spatial and temporal variation in calf
474 survival on the growth of elk populations. *J Wildl Manage* 71:795-803 doi: 10.2193/2005-
475 608

476 Réale D, Festa-Bianchet M (2003) Predator-induced natural selection on temperament in bighorn
477 ewes. *Anim Behav* 65:463-470 doi:10.1006/anbe.2003.2100

478 Réale D, Dingemanse NJ, Kazem AJN, Wright J (2010) Evolutionary and ecological approaches
479 to the study of personality. *Phil Trans R Soc B* 365:3937-3946 doi: 10.1098/rstb.2010.0222

480 Robitaille A, Saucier J-P (1998) Paysages régionaux du Québec méridional. Les publications du
481 Québec, Québec

482 Rognmo A, Markussen KA, Jacobsen E, Grav HJ, Blix AS (1983) Effects of improved nutrition
483 in pregnant reindeer on milk quality, calf birth weight, growth, and mortality. *Rangifer* 3:10-
484 18

485 Sanderson EW, Malanding J, Levy MA, Redford KH, Wannebo AW, Woolmer G (2002) The
486 human footprint and the last of the wild. *BioScience* 52:891-904 doi:
487 [http://dx.doi.org/10.1641/0006-3568\(2002\)052\[0891:THFATL\]2.0.CO;2](http://dx.doi.org/10.1641/0006-3568(2002)052[0891:THFATL]2.0.CO;2)

488 Schlaepfer MA, Runge MC, Sherman PW (2002) Ecological and evolutionary traps. *Trends Ecol*
489 *Evol* 17:474-480 doi: [http://dx.doi.org/10.1016/S0169-5347\(02\)02580-6](http://dx.doi.org/10.1016/S0169-5347(02)02580-6)

490 Seip DR (1991) Predation and caribou populations. *Rangifer* 7:46-52

491 Seip DR, Cichowski DB (1996) Population ecology of caribou in British Columbia. *Rangifer*
492 9:73-80

493 Sih A, Bell A, Johnson JC (2004) Behavioral syndromes: an ecological and evolutionary
494 overview. *Trends Ecol Evol* 19:372-378 doi: <http://dx.doi.org/10.1016/j.tree.2004.04.009>

495 Smith BR, Blumstein DT (2008) Fitness consequences of personality: a meta-analysis. *Behav*
496 *Ecol* 19:448-455 doi: <http://dx.doi.org/10.1093/beheco/arm144>

497 St-Laurent M-H, Dussault C (2012) The reintroduction of boreal caribou as a conservation
498 strategy: A long-term assessment at the southern range limit. *Rangifer* 20:127-138 doi:
499 <http://dx.doi.org/10.7557/2.32.2.2261>

500 Vors LS, Schaefer JA, Pond BA, Rodgers AR, Patterson BR (2007) Woodland caribou
501 extirpation and anthropogenic landscape disturbance in Ontario. *J Wildl Manage* 71:1249-
502 1256 doi: 10.2193/2006-263

503 White RG (1983) Foraging patterns and their multiplier effects on productivity of northern
504 ungulates. *Oikos* 40:377-384

505 Whittington J, Hebblewhite M, DeCesare NJ, Neufeld L, Bradley M, Wilmshurst J, Musiani M
506 (2011) Caribou encounters with wolves increase near roads and trails: a time-to-event
507 approach. *J Appl Ecol* 48:1535-1542 doi: 10.1111/j.1365-2664.2011.02043.x

508 Wikelski M, Cooke SJ (2006) Conservation physiology. *Trends Ecol Evol* 21:38-46 doi:
509 <http://dx.doi.org/10.1016/j.tree.2005.10.018>

510 Williams R, Lusseau D, Hammond PS (2006) Estimating relative energetic costs of human
511 disturbance to killer whale (*Orcinus orca*). Biol Cons 133:301-311 doi:
512 <http://dx.doi.org/10.1016/j.biocon.2006.06.010>

513 Wilson DS (1998) Adaptive individual differences within single populations. Phil Trans Roy Soc
514 B 353:199-205 doi: 10.1098/rstb.1998.0202

515 Wilson DS, Clark AB, Coleman K, Dearstyne T (1994) Shyness and boldness in humans and
516 other animals. Trends Ecol Evol 9:442-446 doi: [http://dx.doi.org/10.1016/0169-](http://dx.doi.org/10.1016/0169-5347(94)90134-1)
517 [5347\(94\)90134-1](http://dx.doi.org/10.1016/0169-5347(94)90134-1)

518 Wittmer HU, McLellan BN, Serrouya R, Apps CD (2007) Changes in landscape composition
519 influence the decline of a threatened woodland caribou population. J Anim Ecol 76:568-
520 579 doi: 10.1111/j.1365-2656.2007.01220.x

Table 1. Candidate models tested to assess the relationship between the habitat selection of female caribou and the fate of their calf (i.e., died from predation or survived) in the boreal forest of Québec, Canada between 2004 and 2011. In candidate models 4 to 12, we added the interaction between calf status and habitat covariates to investigate if habitat selection strategies could explain calf fate. In models 7 to 12, we added three-way interactions to explore if functional responses in habitat selection could explain calf fate.

| Candidate model no. | Variables |
|---------------------|---|
| 1* | Elevation + % Peatlands + % Open lichen woodlands + % Mixed and deciduous |
| 2 | % ≤5-year-old cutovers + % Old 6-20-year-old + % 20-40-year-old cutovers + % Open no regeneration + Road density |
| 3 | Model 1 + Model 2 |
| 4 | Model 1 + Elevation × Calf status + % Peatlands × Calf status + % Open lichen woodland × Calf status + % Mixed and deciduous × Calf status |
| 5 | Model 2 + % ≤5-year-old cutovers × Calf status + % 6-20-year-old cutovers × Calf status + % 20-40-year-old cutovers × Calf status + % Open no regeneration × Calf status + Road density × Calf status |
| 6 | Model 4 + Model 5 |
| 7 | Model 5 + % ≤5-year-old cutovers × % ≤5-year-old cutovers in the calving home range × Calf status |

- 8 Model 5 + % 6-20-year-old cutovers \times % 6-20-year-old cutovers in the calving home range \times Calf status
- 9 Model 5 + % \leq 5-year-old cutovers \times % \leq 5-year-old cutovers in the calving home range \times Calf status + % 6-20-year-old cutovers \times % 6-20-year-old cutovers in the calving home range \times Calf status
- 10 Model 6 + % \leq 5-year-old cutovers \times % \leq 5-year-old cutovers in the calving home range \times Calf status
- 11 Model 6 + % 6-20-year-old cutovers \times % 6-20-year-old cutovers in the calving home range \times Calf status
- 12 Model 6 + % \leq 5-year-old cutovers \times % \leq 5-year-old cutovers in the calving home range \times Calf status + % 6-20-year-old cutovers \times % 6-20-year-old cutovers in the calving home range \times Calf status

*The % of coniferous stands was removed to prevent multicollinearity

Table 2. Survival rate and mortality agent of caribou calves ($n = 30$) monitored during their first 140 days of life from 2009 to 2011 in Saguenay–Lac-St-Jean, Québec, Canada.

| Variable | Year | | | Total |
|-------------------------|-------|-------|-------|-----------|
| | 2009 | 2010 | 2011 | |
| No. of calves | 9 | 11 | 10 | 30 |
| Survival rate | 44.4% | 36.4% | 50.0% | 43.3% |
| Mortality agent (n) | | | | |
| Black bear | 3 | 2 | 4 | 9 (52.9%) |
| Wolf | 0 | 0 | 0 | 0 (0.00%) |
| Unknown predator | 2 | 0 | 1 | 3 (17.6%) |
| Drowning | 0 | 1 | 0 | 1 (5.88%) |
| Unknown* | 0 | 4 | 0 | 4 (23.5%) |

*Mortalities of unknown cause occurred when the carcass was seemingly untouched and there was no evidence of predation.

Table 3. Rankings of the candidate models tested at three spatial scales to assess the relationship between habitat selection by female caribou and the fate of their calf (i.e., died from predation or survived; $n = 57$) in the boreal forest of Québec (Canada) between 2004 and 2011. Candidate models are listed with log-likelihood (LL), numbers of parameters (K), difference in Akaike Information Criterion value versus the best-supported model (Δ AIC), and their relative weight (AICw).

| Candidate model no. | Annual home range scale* | | | | Calving home range scale* | | | | Fine scale | | | |
|------------------------|--------------------------|----|--------------|------|---------------------------|----|--------------|------|------------|----|--------------|------|
| | LL | K | Δ AIC | AICw | LL | K | Δ AIC | AICw | LL | K | Δ AIC | AICw |
| 1 | -38.38 | 5 | 0 | 0.54 | -34.25 | 5 | 0 | 0.73 | -6865 | 5 | 2094 | 0.00 |
| 2 | -37.57 | 6 | 0.38 | 0.44 | -35.04 | 6 | 3.58 | 0.12 | -6772 | 6 | 1910 | 0.00 |
| 3 | -36.85 | 10 | 6.39 | 0.02 | -30.84 | 10 | 3.18 | 0.15 | -6146 | 10 | 666 | 0.00 |
| 4 | | | | | | | | | -6829 | 10 | 2032 | 0.00 |
| 5 | | | | | | | | | -6636 | 12 | 1650 | 0.00 |
| 6 | | | | | | | | | -6004 | 20 | 402 | 0.00 |
| 7 | | | | | | | | | -6509 | 16 | 1404 | 0.00 |
| 8 | | | | | | | | | -6547 | 16 | 1480 | 0.00 |
| 9 | | | | | | | | | -6429 | 20 | 1252 | 0.00 |
| 10 | | | | | | | | | -5873 | 24 | 148 | 0.00 |

| | | | | |
|----|-------|----|-----|------|
| 11 | -5916 | 24 | 234 | 0.00 |
| 12 | -5795 | 28 | 0 | 1.00 |

*At the annual and calving home range scales we only tested candidate models 1, 2 and 3.

Table 4. Coefficient (β) and their 95%CI of the covariables present in the best-supported model at the calving home range (left panel) and local (right panel) scales assessing the link between a female calf 's fate ($n = 57$; i.e., died from predation coded 1, survived coded 0) and the composition of its calving home range (left panel) or its habitat selection strategy (right panel) in the boreal forest of Québec (Canada) between 2004 and 2011.

| Variable | Calving home range scale | | Local scale | |
|------------------------------------|--------------------------|--------------------------|-------------|--------------------------|
| | β | 95%CI [Lower : Upper] | β | 95%CI [Lower : Upper] |
| Intercept | 0.768 | [-3.201 : 4.677] | -4.221 | [-5.063 : -3.745] |
| Elevation (km) | -0.516 | [-5.048 : 4.016] | 8.457 | [7.956 : 9.272] |
| Peatlands (%) | 17.278 | [-9.857 : 44.416] | -8.002 | [-8.172 : -7.814] |
| Open lichen woodland (%) | -34.459 | [-64.054 : -4.865] | 0.729 | [0.673 : 0.809] |
| Mixed and deciduous (%) | -1.360 | [-11.368 : 8.648] | -2.666 | [-2.833 : -2.521] |
| Road density (km/km ²) | | | -5.286 | [-5.414 : -5.169] |
| ≤5-year-old cutovers (%) | | | -7.643 | [-7.786 : -7.457] |
| 6-20-year-old cutovers (%) | | | -5.053 | [-5.405 : -4.680] |
| 20-40-year-old cutovers (%) | | | -4.650 | [-4.709 : -4.544] |

| | | |
|---|--------|--------------------|
| Open no regeneration (%) | -5.498 | [-5.567 : -5.450] |
| Calf alive | -0.340 | [-3.589 : 2.106] |
| ≤5-year-old cutovers in calving home range (%) | -4.934 | [-6.439 : -3.020] |
| 6-20-year-old cutovers in calving home range (%) | -3.953 | [-5.482 : -2.756] |
| Elevation (km) × calf alive | 0.755 | [-2.910 : 5.532] |
| Peatlands (%) × calf alive | -0.104 | [-4.821 : 4.591] |
| Open lichen woodland (%) × calf alive | -3.735 | [-8.732 : 0.375] |
| Mixed and deciduous (%) × calf alive | 4.405 | [0.991 : 7.659] |
| Road density (km/km ²) × calf alive | -9.383 | [-18.472 : -1.962] |
| ≤5-year-old cutovers (%) × calf alive | 4.501 | [1.918 : 7.193] |
| 6-20-year-old cutovers (%) × calf alive | -6.312 | [-21.614 : 4.208] |
| 20-40-year-old cutovers (%) × calf alive | 0.657 | [-1.648 : 3.185] |
| Open no regeneration (%) × calf alive | -1.568 | [-18.892 : 6.212] |
| ≤5-year-old cutovers (%) × ≤5-year-old cutovers in calving home range (%) | 37.696 | [36.699 : 38.481] |
| ≤5-year-old cutovers in calving home range (%) × calf alive | 1.854 | [-1.202 : 3.984] |

| | | |
|--|---------|--------------------|
| 6-20-year-old cutovers (%) × 6-20-year-old cutovers in calving home range (%) | 22.386 | [19.505 : 25.259] |
| 6-20-year-old cutovers in calving home range (%) × calf alive | -1.731 | [-6.958 : 2.637] |
| ≤5-year-old cutovers (%) × ≤5-year-old cutovers in calving home range (%) × calf alive | -19.936 | [-31.583 : -3.178] |
| 6-20-year-old cutovers (%) × 6-20-year-old cutovers in calving home range (%) × calf alive | 20.820 | [-16.650 : 73.138] |

FIGURE LEGENDS

Figure 1. Survival functions of caribou (*Rangifer tarandus caribou*) calves monitored in Saguenay–Lac-St-Jean (2009-2011; $n = 30$) and in Charlevoix (2004-2007; $n = 64$; Pinard et al. 2012) for the first 140 days of life. Detailed information on calf survival for the Charlevoix region can be found in Pinard et al. (2012).

Figure 2. Relative probability of occurrence of female caribou (*Rangifer tarandus caribou*) that did or did not lose their calves to predation ($n=57$), as a function of the proportion of ≤ 5 year-old cutovers within the female calving home range, for three ≤ 5 year-old cutover densities (i.e., 0.0%, 4.6%, 29.1% of 816-m radius circular buffer, corresponding to absent, low and high density of cutovers, respectively).

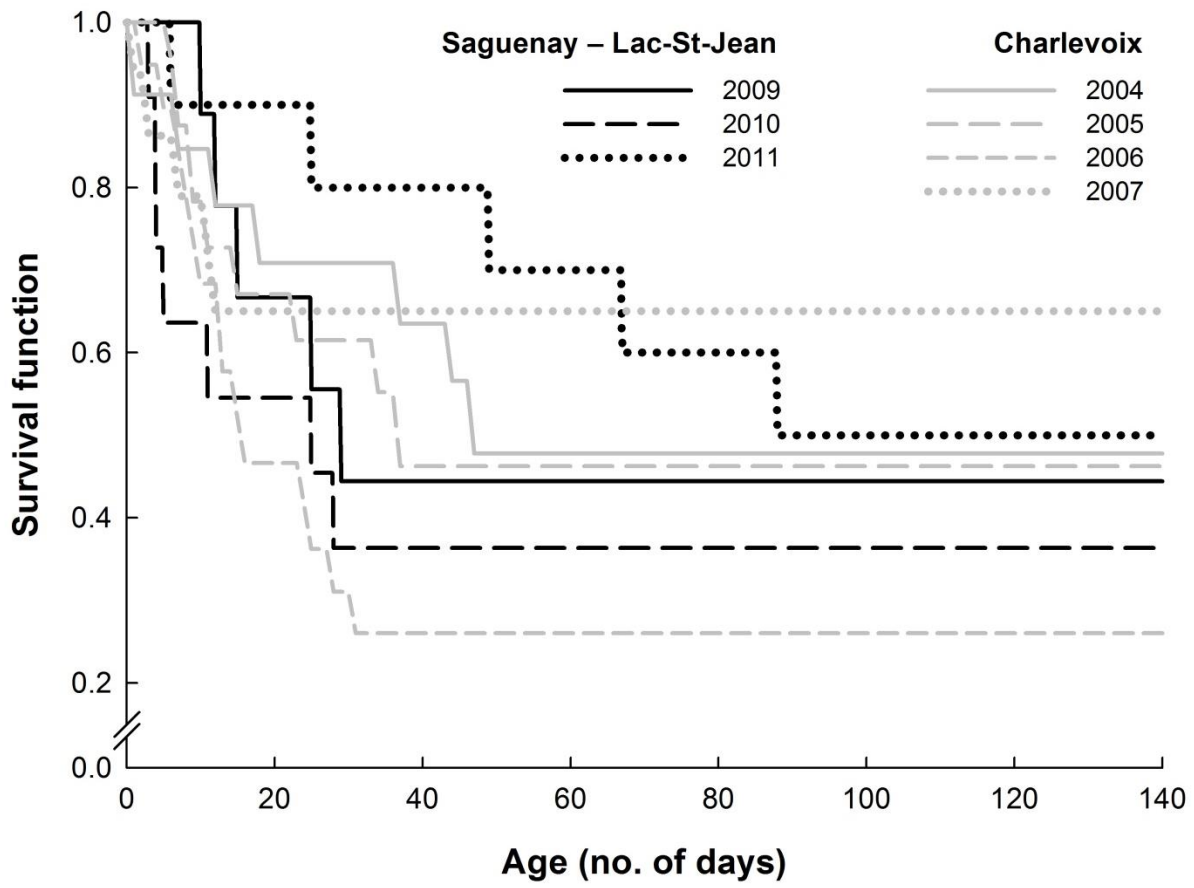


Figure 1.

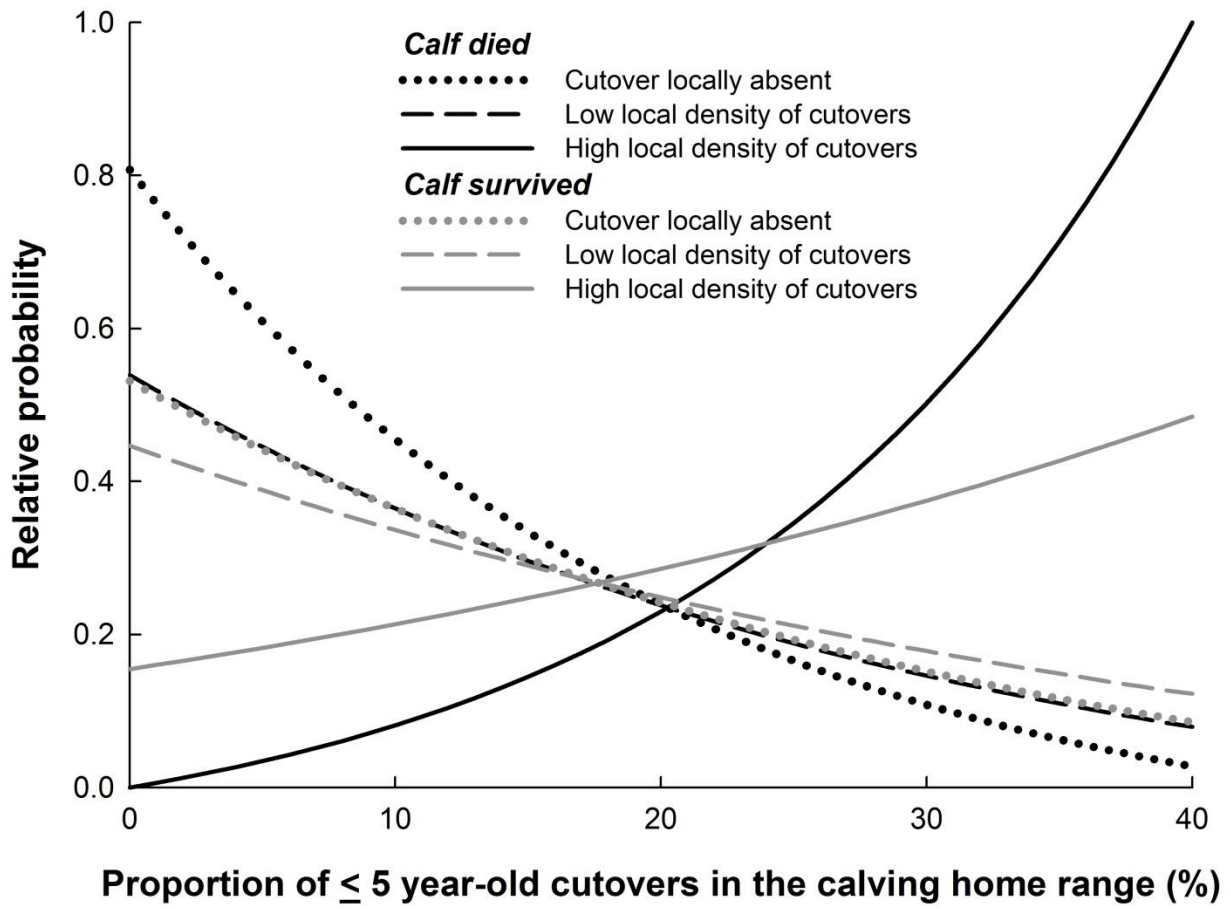


Figure 2.

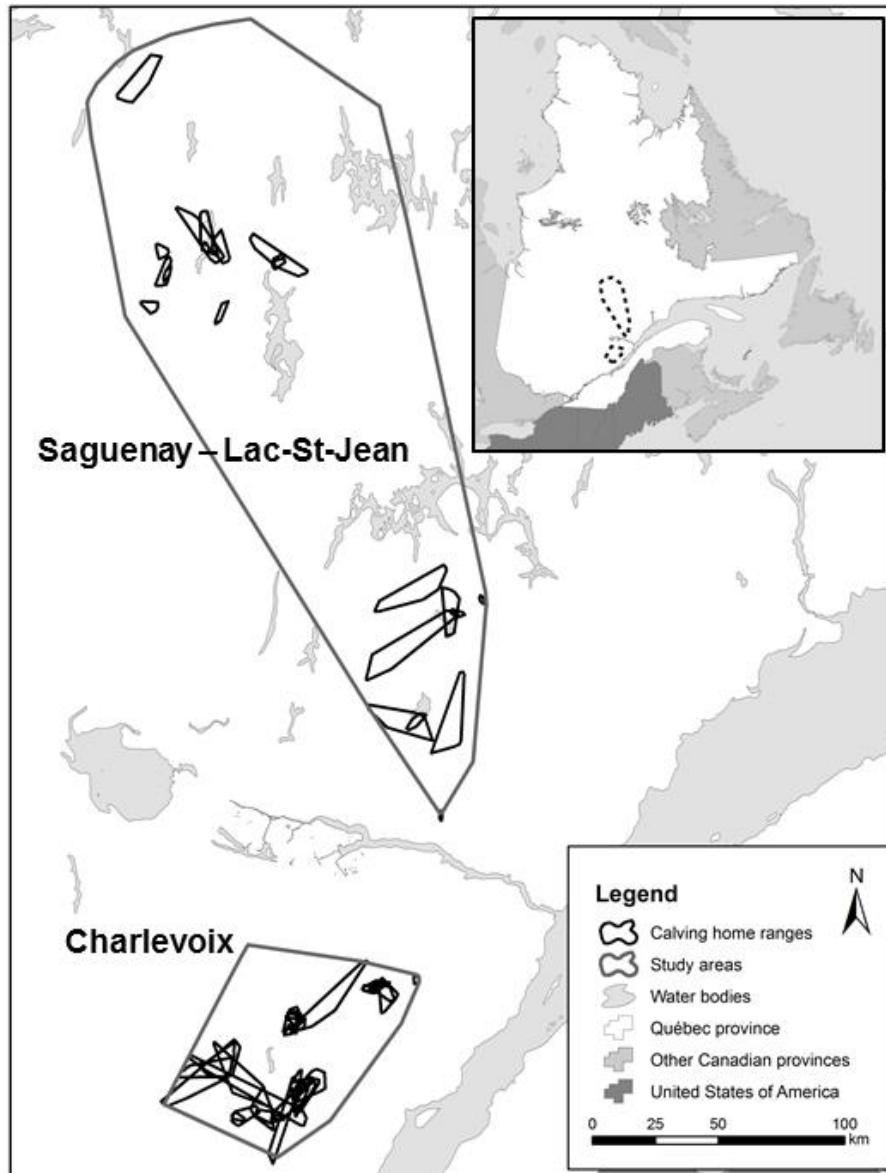


Figure S1. Map of the Saguenay–Lac-St-Jean and Charlevoix study areas and caribou calving home ranges (Québec, Canada).

Article title: Behavioural strategies towards human disturbances explain individual performance in woodland caribou.

Journal: *Oecologia*

Authors: Martin Leclerc, Christian Dussault & Martin-Hugues St-Laurent

Corresponding author: Martin-Hugues St-Laurent, Département de Biologie, Chimie & Géographie, Centre for Northern Studies & Centre for Forest Research, Université du Québec à Rimouski, E-mail: martin-hugues_st-laurent@uqar.ca